

3 **Household slow sand filters operating in continuous and intermittent flows:**
4 **computational fluid dynamics simulation and validation by tracer experiments**

5
6 Liri Yoko Cruz Prieto Hojo^a; Ricardo Vicente de Paula Rezende^c; Sandro Rogério
7 Lautenschlager^b; Lyda Patricia Sabogal-Paz^a

8
9 ^a Department of Hydraulics and Sanitation, São Carlos School of Engineering,
10 University of São Paulo, Trabalhador São-Carlense Avenue, 400, São Paulo 13566-590,
11 Brazil.

12
13 ^b Maringá State University, Technology Center, Department of Civil Engineering,
14 Colombo Avenue, 5790, Maringá 87020900, Paraná, Brazil.

15
16 ^c Maringá State University, Technology Center, Department of Chemical Engineering,
17 Colombo Avenue, 5790, Maringá 87020900, Paraná, Brazil.

18
19 ^a Email: lysaboga@sc.usp.br

20
21 **Abstract**

22 This paper analysed geometries and flows in household slow sand filters (HSSFs)
23 through computational fluid dynamics (CFD) to evaluate the hydrodynamics of filters.
24 Four HSSFs in full scale were studied with a capacity to produce 48 L.d⁻¹ each with
25 diameters of 190 mm and 250 mm and with a filter medium depth of 0.5 meters. The

26 hydrodynamics of the four mathematical models of the HSSFs was validated using the
27 experimental residence time distribution (RTD). The Kruskal Wallis non-parametric test
28 indicates that the results of the experimental RTDs and the simulated RTDs did not
29 show significant differences, therefore the mathematical models represent the physical
30 models, which requires a detailed 3D analysis of the flow in the filters. The results
31 showed that the HSSFs have a flow close to plug flow reactor. Internal recirculation
32 zones were not found, nor short-circuit ones; however, dead zones were verified at the
33 base of the filter with volumes below 3% when compared to the total volume of the
34 filtering and draining layers, not necessary to make changes in geometry. The results of
35 the computational simulation showed that the continuous filters had a smaller velocity
36 variation and the filters with smaller diameter presented a reduction in dead zones when
37 compared to the filters with larger diameters operated in the same flow regime. The
38 study focused on the hydraulic aspects of HSSFs, but it is noteworthy that the choice of
39 the type of operation to be adopted by families that use this type of treatment depends
40 on studies that assess the efficiency of water treatment filters that were not considered in
41 this work.

42

43 **Keywords:** biosand filter, computational fluid dynamics (CFD); drinking water;
44 residence time distribution (RTD); software *ANSYS Fluent*

45

46 **1. Introduction**

47 The ingestion of water contaminated by feces, the lack of collection and
48 treatment of sanitary sewage and inadequate hygiene habits are the causes of about 4
49 billion diarrheal diseases per year, and out of these 1.8 million are fatal (UNEP, 2016).
50 According to Dadonaite et al. (2020), in 2016, it was estimated that 1.55 million deaths

51 were caused by diarrheal diseases worldwide and, in 2017, that number increased to 1.6
52 million; one third of whom were children under the age of five.

53 An inexpensive and efficient way to empower people without access to water,
54 both in urban and rural areas that are not provided with a distribution network, is to use
55 the so-called point-of-use treatment technologies (POU) and, among them, Household
56 Slow Sand Filters (HSSFs) can be highlighted (Sobsey et al., 2008).

57

58 **1.1. Features and functioning of HSSFs**

59 HSSFs were adapted from conventional slow sand filters (SSFs) designed to
60 operate in batch. Both the HSSF and the SSF allow the development of a biological
61 layer on top of the granular medium that helps the treatment (CAWST, 2012; Maciel
62 and Sabogal-Paz, 2020; Andreoli and Sabogal-Paz, 2020).

63 Research carried out on HSSFs showed how easy it is for a user to construct and
64 operate them. These studies focus on evaluating the efficiency (Kennedy et al., 2012;
65 Maciel and Sabogal-Paz, 2020; Terin and Sabogal-Paz, 2019; Sabogal-Paz et al 2020;
66 Andreoli and Sabogal-Paz, 2020).

67 The fundamentals of HSSF contaminant removal are the combination of
68 physical and biological processes. According to Jadhav, Husain and Chavan (2015),
69 physical mechanisms include retention of particles in the pores of the filter media and
70 an adsorption, which allows for organic removal, while biological components include
71 predation, elimination, natural death/inactivation and partial reduction of organic carbon
72 due to the metabolism of the microorganism. Therefore, the HSSF treats water by a
73 combination of physical processes with biological activity present on the surface of the
74 filter layer (*schmutzdecke*).

75 The I-HSSF can operate with a filtration rate of up to $26.4 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$ (Elliot et
76 al., 2006) and the recommended pause period is in the range of 1 to 48 hours (Cawst,
77 2012).

78 The continuous flow filter (C-HSSF) can be operated with a lower filtration rate,
79 up to $9.6 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$ (Maciel and Sabogal-Paz, 2020) and its feeding can be by gravity
80 or direct pumping, if there is a control of the filtration rate (Sabogal-Paz et al., 2020).

81 In the intermittent flow filter (I-HSSF), the water to be treated must be
82 maintained in the filter medium pores between each batch feed, generating the pause
83 period that is important as it allows physical and biological processes to occur to treat
84 water (Sabogal-Paz et al 2020). The study developed by Elliott et al. (2008) showed that
85 the best filter performance occurs when the volume of water to be treated is close to the
86 pore volume.

87 In the study by Maciel and Sabogal-Paz (2020), continuous filters operated at a
88 filtration rate of $0.68 \text{ m}^3 \cdot \text{m}^2 \cdot \text{day}^{-1}$ obtained residence time distribution (RTDs) of 991
89 min. and 876 min. with a turbidity reduction of 89.4% and 83.8%, respectively, and the
90 intermittent filters showed turbidity removal of 91.8 and 90.5%. Terin and Sabogal-Paz
91 (2019) worked with HSSF operating in continuous flow at a filtration rate of 1.22
92 $\text{m}^3 \cdot \text{m}^2 \cdot \text{day}^{-1}$. and obtained a turbidity reduction of 84.39% for the RTD 417 min. and the
93 filter intermittent showed a reduction in turbidity of 81.41% for the 4-hour pause period
94 and 83.78% for the 12-hour pause period.

95 The C-HSSF was classified as a piston flow reactor by Maciel and Sabogal-Paz
96 (2018), Sabogal-Paz et al. (2020), Terin and Sabogal-Paz (2019) and Young-Rojanschi
97 and Madramootoo (2014b), there is a gap in these studies in terms of not analyzing the
98 influence of possible short circuits and dead zones on residence time and quality of

99 water. This question was addressed in the present study using the CFD to analyze the
100 DTR, the type of flow and possible short-circuit influences and dead zones.

101

102 **1.2. Tracer test**

103 The simplest and most direct way to determine the residence time distribution
104 (RTD) is to use a non-reactive tracer, using the stimulus and response technique.
105 Various types of experiments can be used with pulse and/or step type stimuli
106 (Levenspiel, 1999).

107 In real chemical reactors, in general, they tend to deviate from ideality, but for
108 the hydrodynamic characterization of reactors the ideal flows are considered for
109 modelling, which are: Plug Flow Reactor - (PFR) and Continuous Stirred Tank Reactor
110 – (CSTR) (Levenspiel, 1999). In the studies by Andreoli and Sabogal-Paz (2020),
111 Bradley et al. (2011), Elliott et al. (2008), Maciel and Sabogal-Paz (2020), Lynn et al.
112 (2013), Sabogal-Paz et al. (2020) and Terin and Sabogal-Paz (2019), the hydraulic
113 behaviour of the I-HSSF was evaluated and classified as a plug flow reactor. The C-
114 HSSF was also classified as a plug flow reactor by Maciel and Sabogal-Paz (2020),
115 Sabogal-Paz et al. (2020), Terin and Sabogal-Paz (2019), Young-Rojanschi and
116 Madramootoo (2014a).

117 The necessary condition for plug flow reactors is that the residence time must be
118 the same for all fluid elements (Levenspiel, 1999). Thus, the fluid passes through the
119 filter without mixing the fluid from the front and back inlet, that is, without
120 overflowing. It should be noted that in a plug flow reactor, the composition of the fluid
121 varies from point to point along a flow path (Levenspiel, 1999). In this context, Elliott
122 et al. (2008) evaluated an HSSF that showed a minimal effect of dispersion by flow
123 paths through porous media. Therefore, the microbial removal process occurs

124 homogeneously in all water portions that enter the HSSF and advance with the depth in
125 the biological layer (Sabogal-Paz et al., 2020).

126

127 **1.3. Mathematical modelling and simulation of HSSFs**

128 Mathematical models can predict results from the control and operating
129 conditions defined by users to foresee improvements in a filtration process when
130 producing drinking water (Botari and Di Bernardo, 2012). The mathematical models
131 used in the CFD are obtained through differential or integral balances of the conserved
132 properties, that is, they adopt the principles of conservation of mass, momentum and
133 energy.

134 The CFD has become a robust tool in the reactor design and diagnosis process,
135 providing detailed information about the reactors, such as the velocity distribution field,
136 the concentration distribution and energy consumption (Zhang et al., 2007).

137 According to Langergraber et al. (2009), mathematical models were developed
138 and validated with experimental tests, thus they can be used to evaluate and improve
139 existing design criteria. The CFD is normally used in the tracer test simulation and
140 validated with the physical experiment tracer test (Brannock et al., 2010; Furman and
141 Stegowski, 2011; Gharibian et al., 2020; Klusener et al., 2007; Qi et al., 2013; Zhang et
142 al., 2007).

143

144 **1.4. Tracer test with CFD**

145 The comparison of the experimental RTD with the CFD by RTD prediction is
146 essential to ensure an acceptable precision of the mathematical model (Furman and
147 Stegowski, 2011).

148 Regarding the application of the RTD technique to filters, Qi et al. (2013)
149 performed tracer tests with CFD in an upstream sand filter with a length of 2.0 m and
150 width 1.0 m and flow rate of 25.2 L.h⁻¹ and determined the RTD to characterize the flow
151 in filters with two configurations, which are: the first, without a support layer, without a
152 layer of water on the filter layer and with a collector tube along the entire length in the
153 upper portion of the filter layer; and the second, with a support layer, with a layer of
154 water on the filter layer and with a collector tube in part of the upper portion of the filter
155 layer.

156 The CFD simulation, developed by Qi et al. (2013), indicated that the water
157 current lines had an upward and downward sloping direction, and the velocity in the
158 middle of the filter was the highest; about five times greater than the minimum. The
159 comparison of the experimental results with the CFD simulation concludes that the filter
160 that had the support layer, the water slide and partial collector tube eliminated the
161 preferred paths and the stagnation zones compared to the filter that did not have these
162 characteristics. In filters with preferential flow paths, it was proven that the RTD
163 obtained with CFD was lower than the calculated RTD.

164 Although there is research on the hydrodynamics of HSSFs, little has been
165 discussed about the influence of the outlet system on the hydrodynamics of filters
166 (Berbert et al., 2016). The study by Berbert et al. (2016) consisted of analysing the
167 influence of the HSSF output system on the hydrodynamics of the filters and three
168 different models operated in intermittent flow were simulated in the ANSYS-CFX
169 software: the PEU/ UEM model (Berbert et al., 2016), the CAWST version 10
170 (CAWST, 2012) and HydrAid (Kikkawa, 2008).

171 The filters of the CAWST and HydrAid models were simulated with the original
172 geometry that has an outlet system on the side and an outlet configuration with a

173 collector ring inside was simulated. In the PEU/UEM model, four types of localized
174 outlets were simulated: in the center in an upward flow, on the side, at the bottom and a
175 collector ring inside. The study concluded that the PEU/UEM filter that has a collector
176 ring-shaped output with the interface downwards contributes to improving the current
177 lines by creating a better distribution of velocity across the three axes, reducing dead
178 zones and increasing hydraulic residence time (Berbert et al., 2016).

179

180 **2. Materials and methods**

181 **2.1 Experimental**

182 **2.1.1 Construction and operation of HSSFs**

183 Four HSSFs were evaluated, two in continuous flow (C-HSSFs) and two in
184 intermittent flow (I-HSSFs). The filters were built with four layers of the same
185 thickness composed of fine sand, coarse sand, fine gravel and coarse gravel, and above
186 the fine sand there was a water slide, which is presented in detail in the supplementary
187 material. The maximum water level above the filter layer was regulated by a float valve.

188 The C-HSSF1 was a continuous flow filter built in acrylic with a diameter of
189 190 mm and the C-HSSF2, which was also operated in continuous flow, however it was
190 built in PVC with a diameter of 250 mm. The I-HSSF1 operated in an intermittent flow
191 with a 190 mm diameter acrylic body and the I-HSSF2 had the same flow as the
192 previous one, but it was built in PVC with a diameter of 250 mm. Filters with a
193 diameter of 190 mm had a cross-sectional area of 0.028 m^2 and filters with a diameter of
194 250 mm had a cross-sectional area of 0.049 m^2 . The base of the HSSFs consisted of a
195 CAP-type connection, and in the center of that base an outlet tube (collector) was fitted
196 with a tap to capture the filtered water.

197 The walls of C-HSSF1 and I-HSSF1 were covered to prevent algae proliferation
198 inside by sunlight entering through the acrylic material. The filtration rate of continuous
199 filters was $1.69 \pm 0.01 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$ for C-HSSF1 and $0.97 \pm 0.96 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$ for C-
200 HSSF2. On the other hand, the I-HSSF hydraulic head was variable and, consequently,
201 the filtration rate as well. The maximum filtration rate of I-HSSF1 was $3.2 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$
202 shortly after feeding and decreased to zero and the maximum filtration rate of I-HSSF2
203 was $3.94 \text{ m}^3 \cdot \text{m}^{-2} \cdot \text{day}^{-1}$. The filters had a daily production of 48 L each.

204

205 **2.1.2 Physical characterization of materials**

206 The thickness of the water layers (top of fine sand), fine sand, coarse sand, fine
207 gravel and coarse gravel, respectively, were 10 cm, 50 cm, 5 cm, 5 cm and 7.5 cm. The
208 fine sand had an equivalent diameter of 0.17 mm (D_{10}), a uniformity coefficient of 1.73
209 average particle diameter of 2.70×10^{-4} m. The physical parameters used in the study,
210 such as: the thickness of the layers, granulometry, specific mass, the void index and the
211 permeability were the same for the four filters and the standards used in the tests are
212 contained in the supplementary material.

213

214 **2.1.3 Experimental RTD**

215 The disturbance chosen for C-HSSFs and I-HSSFs was of the step type, as
216 reported by Maciel and Sabogal-Paz (2020), Terin and Sabogal-Paz (2019) and
217 Sabogal-Paz et al (2020).

218 Tracer tests on each of the filters were performed in triplicate with a sodium
219 chloride solution with a concentration of $100 \text{ mg} \cdot \text{L}^{-1}$. The variation in the electrical
220 conductivity of the filtered water was measured at the filter outlet, in real time, with a
221 conductivity probe from Venier Software & Technology, USA.

222 The conductivity probe was positioned at the outlet of the tap and inside a bottle.
223 This bottle made it possible to read the samples as the treated water that came out
224 through the tap was collected and temporarily stored in that bottle and renewed each
225 time a new sample of treated water was received.

226 The calibration of the probe allowed to correlate the variation in conductivity
227 with the concentration of the tracer. The probe was connected to a computer with the
228 Logger Lite software (Venier Software & Technology, USA) that allowed us to collect
229 and store the data of the measurements of the tracer concentration over time. The
230 dynamic response time of the conductivity probe was 5 seconds and the data were
231 measured every 60 s in the treated water that was collected at the filter outlet.

232 Data processing was performed using Excel 2013 software (Microsoft®, USA)
233 and OriginPro 8 software (OriginLab, USA). The concentration versus time curve was
234 normalized to obtain the F curve and then derived to obtain the E curve. After each test,
235 the filters were cleaned by introducing tap water until the tracer was completely
236 eliminated.

237 In the C-HSSFs, the tracer tests were performed with the same filtration rate
238 used during the operation of the filters. The procedure resulted in the concentration
239 versus time curve (C vs. t). These experimental curves were adjusted based on the three
240 mathematical models of dispersion: model of continuous mixing tank reactors (N-
241 CSTRs) and the models of low and high dispersion, as reported by Levenspiel (1999).

242 In the I-HSSFs tracer tests, the disturbance was performed with the first feeding
243 with 100 mg.L^{-1} salt solution and then with tap water until the tracer concentration was
244 close to zero. The filtration rate decreased to zero when the hydraulic load reached the
245 minimum water level, at which time new feeding was performed. The volume of each

246 feed corresponded to the volume of voids in the filter. In total, four 10 L feeds were
247 used in I-HSSF1 and four 15 L feeds in I-HSSF2.

248 Afterwards, the Modified Morris Dispersion Index (MMDI) proposed by Lynn
249 et al. (2013) was calculated, which is a modification of the Morrill Dispersion Index
250 (MDI) according to Tchobanoglous et al. (2003). The MDI index ranges from 1.0 (ideal
251 piston flow reactors - PFR) to 22 (ideal complete mix reactors - CSTR) according to
252 Tchobanoglous et al. (2003).

253 The theoretical RTD was calculated considering the ratio between the volume of
254 water stored by C-HSSF1 and C-HSSF2 of 10L and 15L, respectively, and the daily
255 flow of 48 L for both.

256

257 **2.2. Mathematical modelling and simulation of HSSFs**

258 The simulations of the HSSFs solved the mass conservation and momentum
259 conservation equations using the ANSYS Fluent software (version 2020 R2) based on
260 the finite volume method, where the differential equations are integrated in each mesh
261 control volume, generating a set of discrete algebraic equations. The resulting linear
262 system was then solved numerically in an iterative way until the solution was reached
263 with previously determined precision.

264 The three-dimensional equations of the conservation of mass and amount of
265 movement are given by Eq. 1 (all domains - continuity equation) and Eq. 2 (fluid
266 domain - Navier-Stokes equation) (Navier, 1823; Stokes, 1845). In the filtering and
267 support layers of the HSSFs, the term source of moment (S_i) was added to the fluid flow
268 equations in Eq. 3 (Porous Domain, simple homogeneous porous media of Darcy-
269 Forchheimer Equation) (Darcy, 1856; Forchheimer, 1901a, 1901b). The term source of
270 moment consists of two parts: loss of viscosity (i.e., Darcy's Law in porous media, the

271 first term on the right side of Eq. 3), and inertial loss (i.e., the second term on the right
 272 side of Eq. 3).

$$273 \quad \nabla \cdot (\rho u) = 0 \quad (1)$$

$$\nabla \cdot (\rho u) = 0$$

$$274 \quad \nabla \cdot (\rho u u) = -\nabla p + \nabla \cdot \left(\mu (\nabla u + \nabla u^T) - \frac{2}{3} \delta \nabla \cdot u \right) \quad (2)$$

$$\nabla \cdot (\rho u u) = -\nabla p + \nabla \cdot \left(\mu (\nabla u + \nabla u^T) - \frac{2}{3} \delta \nabla \cdot u \right)$$

$$275 \quad S_i = - \left(\frac{\mu}{\alpha} u_i + C_2 \frac{1}{2} \rho |u| u_i \right) \quad (3)$$

$$S_i = - \left(\frac{\mu}{\alpha} u_i + C_2 \frac{1}{2} \rho |u| u_i \right)$$

276

277 Where, $u \in \mathbb{R}^3$ is the velocity vector; ρ is the density of the fluid (water) [kg.m⁻³];
 278 p is the pressure [atm]; μ is the viscosity [m².s⁻¹]; where, S_i is the source term for i in
 279 (x , y or z) in the moment equation, α is the permeability of the porous medium [cm.s⁻¹],
 280 $|u|$ is the magnitude of the velocity [m.s⁻¹], and D and C are the diagonal matrices
 281 with $1/\alpha$ and C_2 , respectively.

282 Eq. 3 contributes to the pressure gradient in the porous cell, creating a pressure
 283 reduction that is proportional to the velocity of the fluid in the cell (ANSYS, 2013). The
 284 parameters of viscous resistance ($1/\alpha$) and inertial resistance (C_2) used for porous media
 285 are calculated by Eq. 4 and Eq. 5 of the Ergun Equation (Ergun, 1952).

$$286 \quad \alpha = \frac{D_p^2}{150} \cdot \frac{\varepsilon^3}{(1 - \varepsilon)^2} \quad (4)$$

$$\alpha = \frac{D_p^2}{150} \cdot \frac{\varepsilon^3}{(1 - \varepsilon)^2}$$

$$287 \quad C_2 = \frac{3,5}{D_p} \cdot \frac{(1 - \varepsilon)}{\varepsilon^3} \quad (5)$$

$$C_2 = \frac{3,5}{D_p} \cdot \frac{(1 - \varepsilon)}{\varepsilon^3}$$

288 Where: $1/\alpha$ – Viscous resistance [m^2]; D_p - Average particle diameter [m]; ε –
 289 Index of voids [$\text{kg}\cdot\text{m}^{-3}\cdot\text{kg}^{-1}\cdot\text{m}^3$]; C_2 – Inertial resistance [m^{-1}]; ε – Fraction of voids
 290 [$\text{kg}\cdot\text{m}^{-3}\cdot\text{kg}^{-1}\cdot\text{m}^3$].

291 In laminar flows through porous media, the pressure drop is typically
 292 proportional to the velocity and the constant C_2 can be considered zero (ANSYS, 2013).
 293 Ignoring convective acceleration and diffusion, the middle model is reduced to Darcy's
 294 Law (Darcy, 1856), according to Eq. 6.

295 (6)

$$\nabla p = -\frac{\mu}{\alpha} u$$

296 Inertial losses in a porous medium, at high flow velocities, can be corrected by
 297 the constant C_2 of Eq. 3. This constant can be seen as a loss coefficient per unit length
 298 along the flow direction, thus allowing the drop of pressure, specified as a function.

299 If the permeability term can be eliminated, only the term of inertial loss can be
 300 used, producing the simplified form of the porous medium equation (Eq. 3) by Eq. 7 in
 301 the x, y and z directions (ANSYS, 2013).

302 (7)

$$\nabla p = -\sum_{j=1}^3 C_{2ij} \left(\frac{1}{2} \rho \mathbf{u} |\mathbf{u}| \right)$$

303

304 The physical domain to be studied in the HSSFs is delineated by the walls of the
 305 filter and its interior, composed of layers of water, filter material and the support layer.
 306 The physical properties of the constituent materials of the physical domain, such as
 307 viscosity, specific mass, pressure, porosity and hydraulic conductivity (or permeability),
 308 were obtained in laboratory tests (supplementary material). The viscosity used was
 309 $1.003 \times 10^{-3} \text{ kg}\cdot\text{m}^{-1}\cdot\text{s}^{-1}$ and the density was $9.982 \times 10^2 \text{ kg}\cdot\text{m}^{-3}$.

310 Thus, the boundary conditions of entry and exit were established for each of the
311 constituent layers of the HSSF and the behaviour of the solution of the equations as a
312 function of time was specified.

313 The CFD simulation was developed using Ansys 2020 R2 software from
314 Ansys® and its subcomponents and a model was generated for each filter. Initially, the
315 geometry of each HSSF was created using Autodesk® AutoCAD 2010®, and then the
316 mesh was generated using Ansys 2020 R2 Meshing (Ansys®) with different refinements.
317 Then, the control conditions were defined to run the simulation in Ansys 2020 R2
318 Fluent (Ansys®) and the results were visualized in Ansys 2020 R2 CFD Post (Ansys®).

319 The models were simulated on a computer with an Intel® Core™ i7-7500U
320 CPU @ 2.70 GHz 2.90 GHz, 64-bit operating system, 8 GB installed memory (RAM), 1
321 TB HD, operating on Microsoft's Windows 10 platform®.

322 The results generated the velocity profiles and the flow current lines inside the
323 filters. At the end of the simulation, the verification and validation of the mathematical
324 models were performed based on the results obtained in the experimental tracer tests.

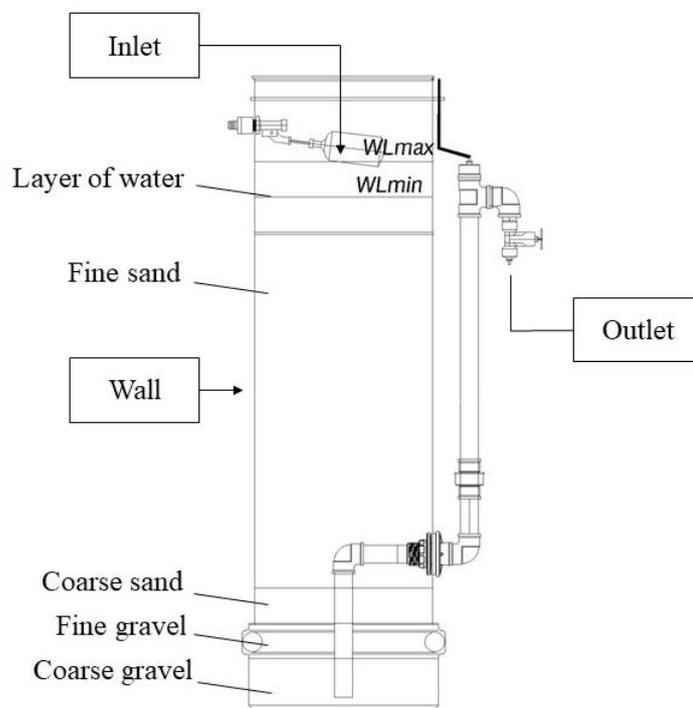
325 In the present study, two simulations were carried out for each HSSF: the first of
326 the flow, which was carried out in a permanent regime; and the second from RTD,
327 performed in a transient regime (passive scalar).

328 In the four mathematical models of the HSSFs, to facilitate modelling, the
329 following criteria were adopted: i) single-phase flow in three dimensions; ii) Newtonian
330 fluid at steady state; iii) isothermal flow, laminar flow, incompressible flow and flow in
331 porous medium in the filtering and draining layers; and iv) homogeneous and specific
332 porosity and permeability for each of the filtering and draining layers; as reported by
333 Berbert et al. (2016).

334 The boundary conditions used for the permanent regime, Navier-Stokes, were:
 335 i) Inlet: prescribed velocity; ii) Outlet: prescribed pressure; iii) Wall condition: non-slip.
 336 The boundary conditions for the transient, passive scalar regime were: i) Inlet:
 337 prescribed concentration of 100; ii) Outlet: null gradient; iii) Wall condition:
 338 impermeable (null normal gradient).

339 The boundary conditions used for the permanent regime, Navier-Stokes, were: i)
 340 Inlet: prescribed velocity; ii) Outlet: prescribed pressure; iii) Wall condition: non-slip.
 341 The boundary conditions for the transient regime, passive scalar, were: i) Inlet:
 342 prescribed concentration of 100; ii) Outlet: null gradient; iii) Wall condition:
 343 impermeable (null normal gradient).

344 In Figure 1, the inlet, outlet and wall of the filter are indicated. The fluid velocity
 345 at the entrance of the C-HSSFs was calculated by the ratio between the flow (48 L.day⁻¹)
 346 and the area of the entrance section.



347
 348 Figure 1. Schematic of the C-HSSFs indicating the inlet, the outlet and the filter wall for
 349 modelling purposes

350 In the I-HSSFs, the fluid velocity at the inlet was obtained from the average
351 velocity calculated in the RTD experimental test. The average flow was obtained by the
352 ratio between the volume of water and the total time of this test and then the velocity
353 was calculated by the ratio between this flow and the area of the inlet section.

354 A time step was selected that guaranteed a Courant Number (Co) or CFL
355 criterion (Courant, Friedrichs and Lewy, 1967), less than 1, according to Eq. 8.

356 (8)

$$Co = \frac{U \cdot \Delta t}{\Delta}$$

357 Where parameter U is the local velocity [$m \cdot s^{-1}$], Δt is the time step [s] and Δ is
358 the local length scale characteristic of the finite volumes of the mesh [m].

359 The Courant Number is a local parameter of the fluid, dimensionless domain,
360 which represents the ratio between the characteristic physical time for the information to
361 propagate the U velocity from one node in the mesh to the next, and the discrete time
362 step used in the simulation. The use of longer time steps ($Co > 1$) locally implies a
363 violation of the transport physics of the property under analysis, which almost always
364 results in numerical oscillations and/or divergence of the numerical solution.

365 After constructing the meshes with different refinements (variation in the number of
366 finite elements), a mesh independence test was carried out. In this test, the velocity
367 obtained for each mesh was analysed to determine from which degree of refinement the
368 calculated values for velocity became independent of the mesh. Therefore, the minimum
369 volume of the finite element (ΔV [m^3]) was obtained from the meshes generated in the
370 CFD, adopting the Courant Number (Co) of 0.9, according to Eq. 8 and Eq. 9.

371 (9)

$$\Delta = \sqrt[3]{\min(\Delta V)}$$

372 The condition used to verify the independence of the mesh was that the velocity
 373 in the filter outlet, located at the centroid of the end of the outlet tube, did not present
 374 variations greater than 5% of the real velocity.

375

376 **2.2.1 Numerical method**

377 Two geometries were created corresponding to each filter diameter (190 mm and
 378 250 mm) and it is not necessary to construct four geometries (one per filter), since the
 379 flow characteristic (continuous or intermittent) depends only on the boundary
 380 conditions inserted in the simulation. The elaborated geometries followed the structure
 381 shown in Figure 1.

382 Table 1 shows the numerical parameters used to simulate the filters and the RTD
 383 simulations. The reference pressure adopted was 1 atm and the relative static pressure
 384 was 0 atm, considering that the experiment was carried out close to sea level and the
 385 flow was free. In the simulations carried out in a transient regime, that is, in the RTD
 386 simulations, the following solution methods were used: i) Transient formulation of “first
 387 implicit order”, ii) SIMPLE coupling scheme.

388

389 Table 1 – Numerical parameters for filter simulation and RTDs simulation

Parameters	C-HSSF1	C-HSSF2	I-HSSF1	I-HSSF2
Reference pressure [atm.]	1	1	1	1
Relative static pressure [atm.]	0	0	0	0
Entrance section area [m ²]	0.0283	0.0491	0.0283	0.0491
Inlet fluid velocity [m.s ⁻¹]	1.96 x10 ⁻⁵	1.13 x10 ⁻⁵	4.10 x10 ⁻⁵	5.66 x10 ⁻⁵
Relative outlet pressure [atm.]	0	0	0	0
Passive scalar value	100	100	100	0

Time step (s)	23	56	600	600		
Number of time steps	1566	1072	13	30	12	24
Maximum number of iterations	15	15	15	15		
Total computational time (h)	5	8	4	4		

Notes: C-HSSF1 - continuous filter with a diameter of 190 mm; C-HSSF2: continuous filter with a diameter of 250 mm; I-HSSF1: intermittent filter with a diameter of 190 mm; and I-HSSF2: intermittent filter with a diameter of 250 mm.

390

391 In I-HSSF1, a volume of 30 L was used in a time of 430 min and in I-HSSF2, 60
392 L was applied in 360 min, obtained in the experimental tracer tests. The relative
393 pressure at the outlet of 0 atm was used, as the flow depended only on the variation of
394 the water level that preceded the layer of fine sand, that is, the flow was free.

395 The calculated values of viscous resistance ($1/\alpha$) and inertial resistance (C_2) were
396 obtained based on the mean particle diameter ((D_p)) and the void index (ϵ) values
397 obtained in a laboratory test are contained in the supplementary material.

398 In the simulations carried out in the HSSFs steady state, the following solution
399 methods were used: i) Pressure-velocity coupling scheme: SIMPLE (Semi-Implicit
400 Method for Pressure Linked Equations); ii) spatial discretization of gradient “cells based
401 on least squares”; iii) “second order” pressure spatial discretization; and iv) spatial
402 discretization of “second-order upwind” momentum. The computational time for each
403 of the simulations was approximately 2 hours and 400 iterations were used. The results
404 of the simulations residues were less than 1×10^{-5} .

405 The dead zones were quantified by occupied volumes between velocities greater
406 than 0 and less than $1.05 \times 10^{-5} \text{ m.s}^{-1}$ for C-HSSF1, $6.90 \times 10^{-6} \text{ m.s}^{-1}$ for C-HSSF2,
407 $2.32 \times 10^{-5} \text{ m.s}^{-1}$ for the I-HSSF1 and $2.22 \times 10^{-5} \text{ m.s}^{-1}$ for the I-HSSF2. These velocities

408 were determined punctually in the velocity contour plane where there was no formation
409 of current lines. After defining this criterion and delineating the dead zone, its volume
410 can be calculated in relation to the sum of the volumes of the filtering and draining
411 layers. The images of the current lines and the velocity contour plane are shown in the
412 supplementary material.

413 **2.2.2 RTD simulation**

414 The determination of the residence time distribution with CFD used the passive
415 scalar equation (TALVY et al., 2011) for a transient regime as a solution for using a
416 tracer concentration in the model.

417 The transport of a passive scalar refers to the convection-diffusion of a scalar
418 field within the fluid in motion without significant changes in the properties of the fluid
419 under analysis, such as in the velocity field, pressure and temperature, hence the term
420 passive. Eq. 10 was used to calculate the passive scalar, and thus solve the tracer test in
421 the model (ANSYS, 2013).

422 (10)

$$\frac{\partial \rho C}{\partial t} + \nabla \cdot (\rho UC) = \nabla \cdot (D_{disp} \nabla C)$$

423 Where: ρ is the density of the fluid in kg.m^{-3} ; C is the tracer concentration in
424 mg.L^{-1} ; U is the velocity vector in m.s^{-1} ; and D_{disp} is the dispersion coefficient given in
425 $\text{m}^2.\text{s}^{-1}$.

426 In this equation, the first term on the left side of the equality corresponds to the
427 concentration of solute in the solvent over time, the second term corresponds to the
428 advective term of concentration carried by the velocity field; and the one on the right
429 side of equality, the term corresponds to the diffusive term of concentration.

430 The ANSYS Fluent software solves the transport equation for a passive scalar
431 defined by the user as User-defined scalar (UDS) in the same way as it solves the

432 transport equation for a scalar such as mass fraction (ANSYS, 2013). For transient
433 simulations, the governing equations were discretized in space and time.

434 In the CFD tracer test, the diffusion coefficient at 25 ° C of NaCl corresponding
435 to $1.612 \times 10^{-9} \text{ m}^2 \cdot \text{s}^{-1}$ with diffusivity of $1.609 \times 10^{-6} \text{ kg} \cdot \text{m}^{-1} \cdot \text{s}^{-1}$ was used, as established by
436 Robinson and Stoke (1955). In the tests of C-HSSFs and I-HSSFs, the dimensionless
437 scalar value of 100 was used to refer to the $100 \text{ mg} \cdot \text{L}^{-1}$ salt solution and 0 for the other
438 tap water feeds. The outgoing concentration of the passive scalar was obtained by the
439 integral of the outgoing section area.

440

441 **2.3. Model validation and verification**

442 Furman & Stegowski (2011) suggested the following steps to verify the CFD
443 results: i) Choose the CFD model appropriately from the physical point of view of the
444 problem; ii) obtain the velocity field and the RTD simulation with CFD; iii) evaluate the
445 deviation of the results of the experimental RTD and RTD predicted with CFD; and iv)
446 analyse the velocity fields if they are different simulated geometries.

447 The article presents the flow simulations applied in HSSFs using the Ansys
448 Fluent 2020 R2 to analyse the influence of continuous (C-HSSF) and intermittent (I-
449 HSSF) feeds and geometric shape changes on the hydrodynamic behaviour of the filter.

450

451 **2.4. Statistical analysis**

452 A priori the data sets were assessed for normality by the Shapiro-Wilk test,
453 which allowed us to classify each data set as: data with normal distribution (Shapiro-
454 Wilk test $p > 0.05$) in which parametric tests are applied or data with non-normal
455 distribution (Shapiro-Wilk test $p < 0.05$) in which non-parametric tests are applied.

456 The statistical analysis used to validate experimental and computational RTD
457 tests was performed using the Kruskal-Wallis non-parametric test (95% confidence
458 limit) followed by the Mann-Whitney pairwise test, indicated when comparing three or
459 more groups, to perform the tests, the free access software PAST 4.03
460 (PAlaeontological STatistics) was used.

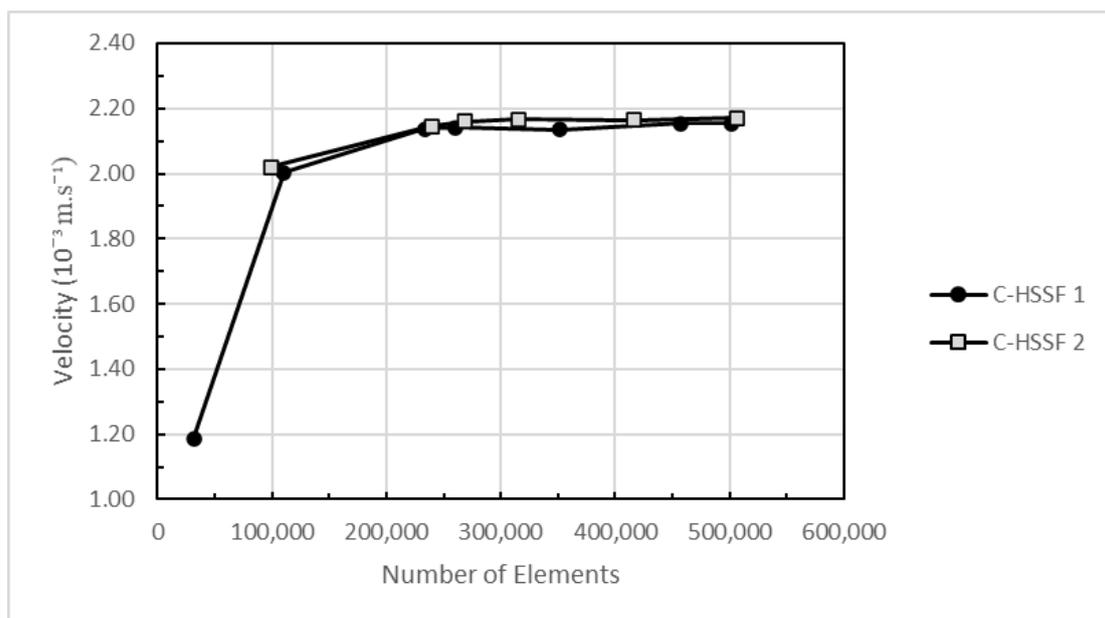
461

462 3. Results and discussion

463 3.1. Numerical Results

464 In steady state simulations, the outlet velocities were obtained, located in the
465 centre of the outlet tube, that is, in the centroid of the end of the outlet tube. The
466 behaviour of this velocity was analysed as a function of the control volumes, as shown
467 in Figure 2. It can be seen that the minimum value required for variations not greater
468 than 5% for the C-HSSF1 was 233,264 elements (Figure 3A) and for C-HSSF2 it was
469 268,474 elements (Figure 3B).

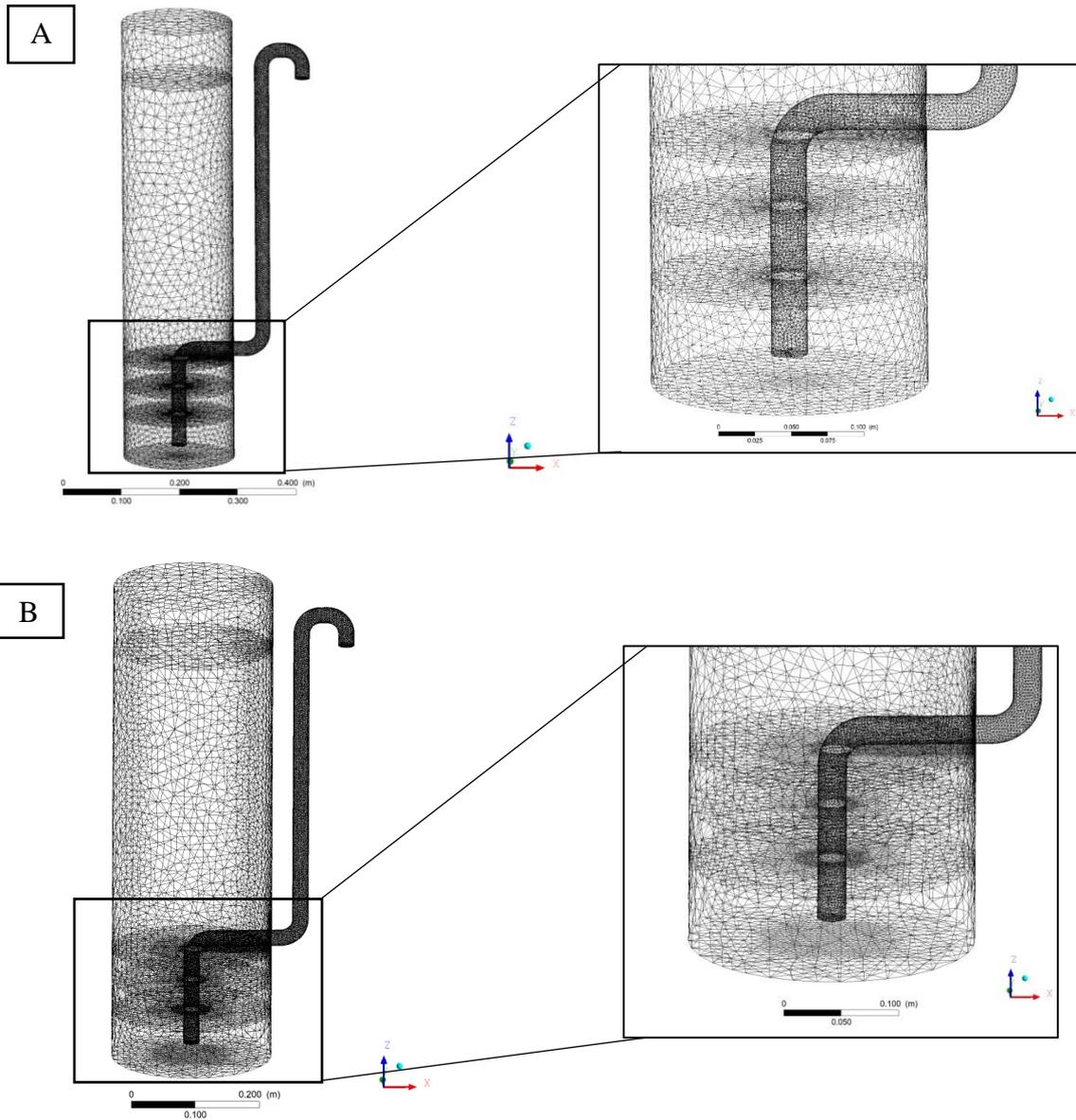
470



471

472 Figure 2 – Velocity as a function of the number of elements of the numerical grid
473 obtained in the mesh independence test of C-HSSF1 and C-HSSF2 using Ansys®
474 Software 19.2.

475



476 Figure 3 – Meshes obtained for C-HSSF1 and C-HSSF2.

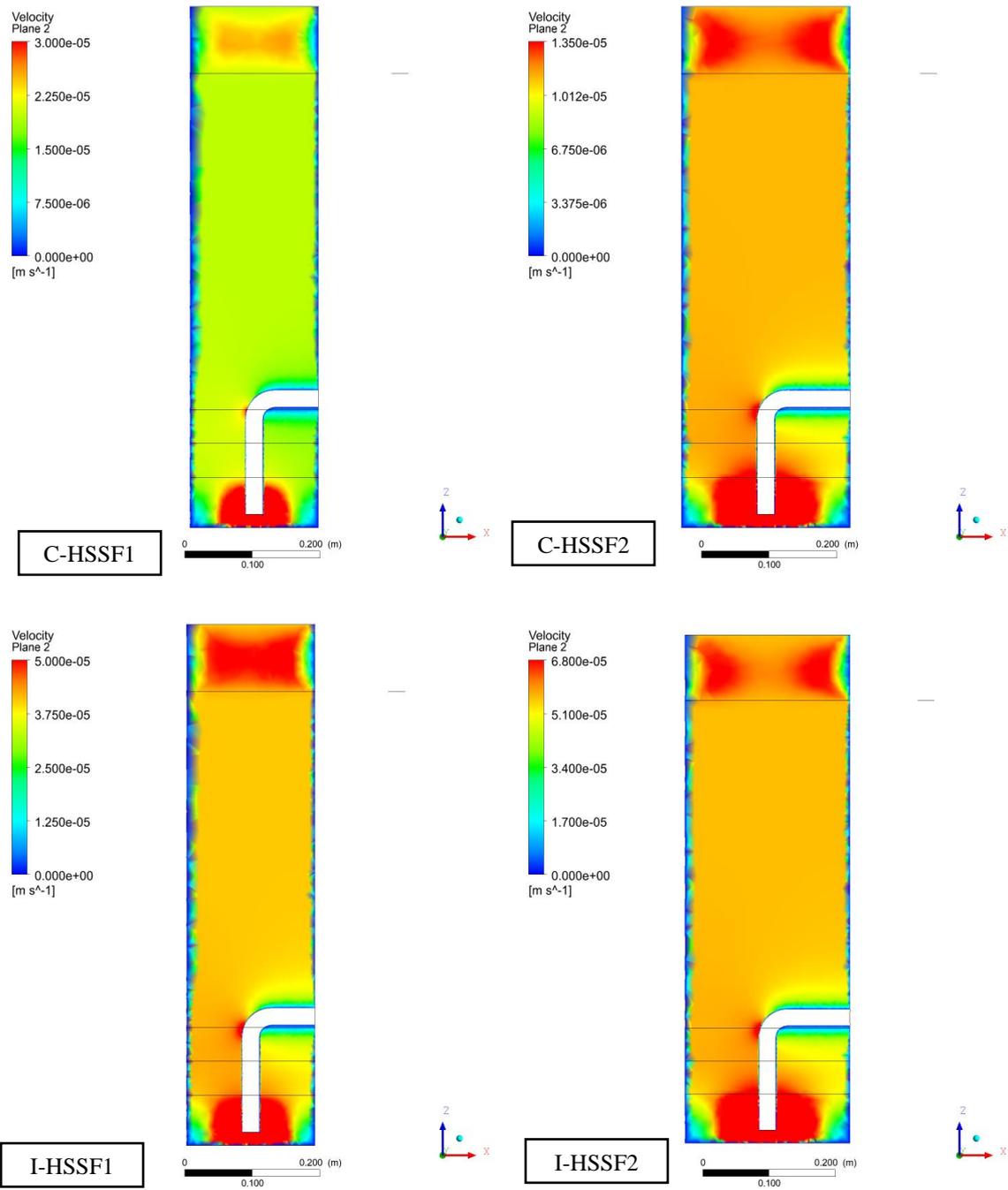
477 Legend: Figure 3A (C-HSSF1 with 233,264 elements of tetrahedral type); and Figure
478 3B (C-HSSF2 with 268,474 elements of tetrahedral type).

479

480 According to the results, the current lines showed symmetry in relation to the
481 central axis, the existence of a dead zone near the base of the HSSF in the coarse gravel
482 layer and the absence of a short circuit. Similar behaviour was described by Berbert et
483 al. (2016) who found the symmetry and dead zone in HSSF, without obtaining the
484 volume of the dead zone, with the same output configuration used in the present study.

485 In the simulated HSSFs, the locations of the dead zone were defined, and the
486 volumes were obtained. The percentage results of the volume of the dead zones in
487 relation to the sum of the volumes of the filtering and draining layers were: C-HSSF1
488 2.19%, C-HSSF2 2.99%, I-HSSF1 2.37% and I-HSSF2 2.60%. It was found that the
489 values were low, below 3%, and the behaviour of the current lines in the HSSFs were
490 similar, regardless of the geometry or type of flow. It was also found that for the same
491 type of flow regime, whether in continuous or intermittent flow, the increase in diameter
492 implied an increase in the volume of the dead zone. The results of the dead zone
493 volumes are shown in the supplementary material.

494 The velocity profiles in the XZ plane at the centre of the HSSFs were visually
495 similar, as seen in Figure 4. The velocity scale was previously defined, and the outlet
496 tube was removed from the visualization for a better analysis of the velocity variation
497 inside the filters. It was found that in the water layer the velocity was higher than when
498 it reached the filtering and draining layers, and then increased in the coarse gravel layer
499 before reaching the water outlet.



500 Figure 4 – Velocity profiles on the XZ plane located in the center of the HSSFs

501 indicating the maximum and minimum velocities

502

503 In the simulations, the velocities were sampled in the centroid of the filter,

504 located in the layer of fine sand, and the velocities of the real experiments were

505 compatible (Table 2), which indicates the representativeness of the mathematical
 506 models in relation to the HSSFs built in real scale.

507

508 Table 2 - Velocity of experimental and simulated HSSFs

HSSF	Velocity (m.s ⁻¹)		Difference (%)
	Experimental	Simulation	
C-HSSF1	1.959x10 ⁻⁵	1.956x10 ⁻⁵	-0.15
C-HSSF2	1.132x10 ⁻⁵	1.129x10 ⁻⁵	-0.27
I-HSSF1	4.10x10 ⁻⁵	4.092x10 ⁻⁵	-2.44
I-HSSF2	5.66x10 ⁻⁵	5.648x10 ⁻⁵	-2.83

509

510 The velocity profiles obtained for the HSSFs, as shown in Figure 5, show that at
 511 the initial points on the water slide the velocity is higher and then there is a reduction
 512 when it starts to percolate through the filter layer. The velocity remains constant until
 513 reaching the last layers, when there is an increase close to the water outlet/treated water
 514 collection region and the last point shows the velocity close to the wall at the bottom of
 515 the filter. The increase in velocity near the water outlet corroborates the observations of
 516 Berbert et al. (2016).

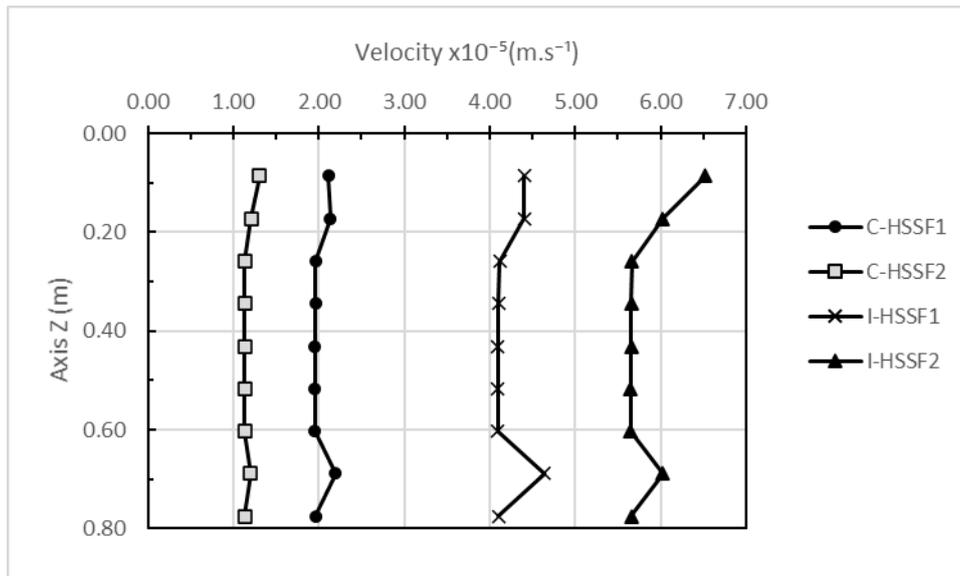


Figure 5 – Velocity profiles of HSSFs.

517

518

519

520

521

522

523

524

525

526

527

528

529

530

531

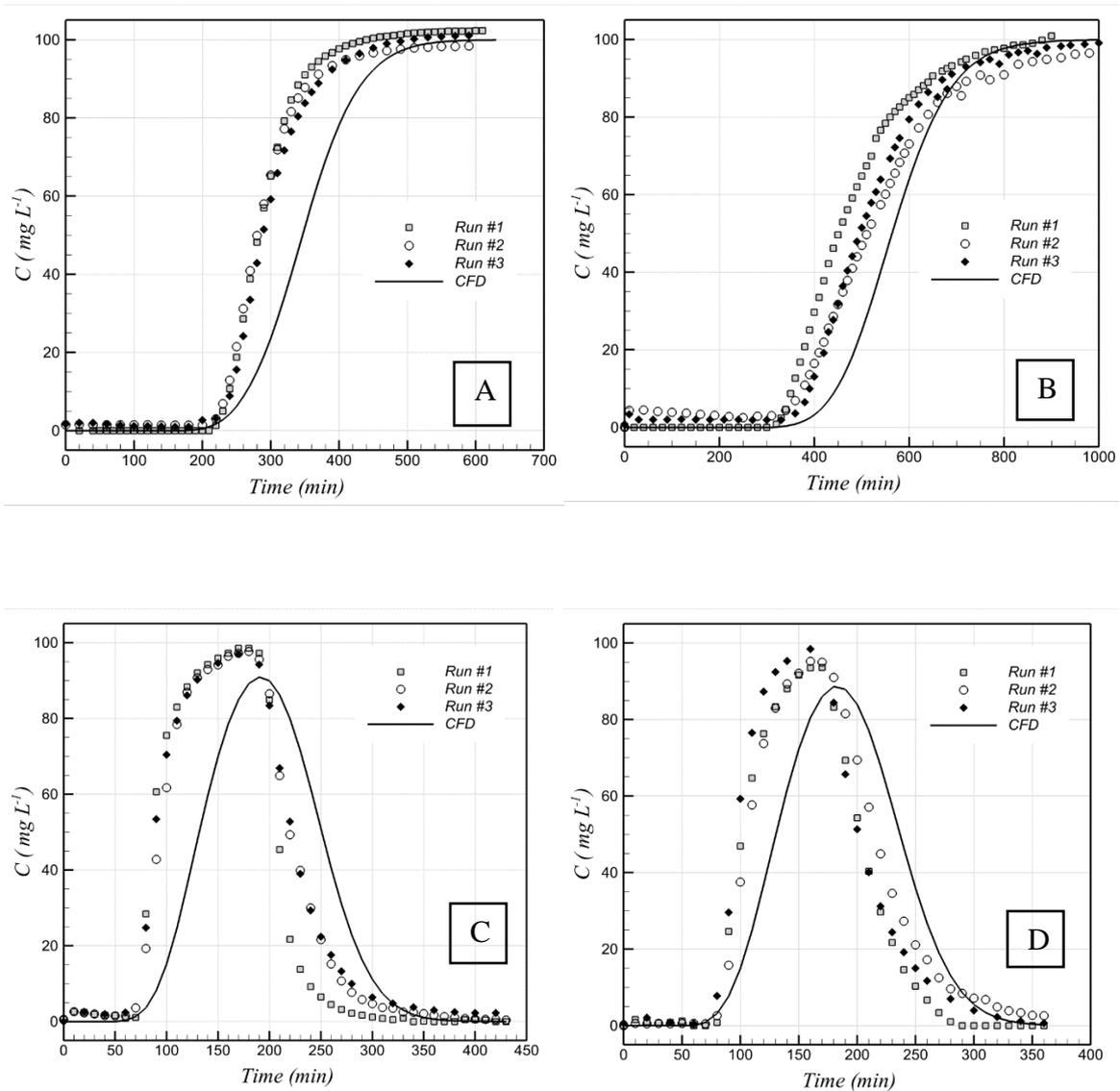
532

533

The analysis of Figures 4 and 5 allows us to highlight two important points. The first point shows that despite the velocity profiles being visually similar, as can be seen in Figure 4, it is observed in Figure 5 that the I-HSSFs present a greater velocity variation than the C-HSSFs at the beginning of the filtering layer. This characteristic can negatively affect the development of the biological layer, which needs an environment in balance for its development and, consequently, reduce the treatment efficiency in filters operated in intermittent flow. The second point shows that this velocity variation is still greater in the C-HSSFs near the base and the outlet collector tube, which indicates that each of the filters needs an individual determination of the limiting velocity to define the dead zone, or that is, the same dead zone delineated velocity cannot be used for the four filters.

3.2. Comparison between experimental and simulated RTD (CFD)

534 Table 2 shows that the velocities obtained in the simulation were close to and
535 below the experimental velocities and the analysis of the results obtained with the
536 experimental tracer tests and the CFD tracer test, presented in Figure 6, can classify the
537 HSSFs as a plug flow reactor.
538



539 Figure 6 – Experimental and computational RTD of HSSFs.

540 Notes: A - C-HSSF1; B - C-HSSF2; C - I-HSSF1; and D - I-HSSF2.

541

542 The results obtained in the I-HSSFs corroborate the behaviours described by
543 Andreoli and Sabogal-Paz (2020), Bradley et al. (2011), Elliott et al. (2008), Maciel and
544 Sabogal-Paz (2020) and Terin and Sabogal-Paz (2019), who characterized the HSSF as
545 a plug flow reactor.

546 According to the adaptation of the MDI to the mMDI proposed by Lynn et al.
547 (2013), they presented mMDI results of 2.92. In the I-HSSFs of Andreoli and Sabogal-
548 Paz (2020), the MDI values were 2.19 and 2.21 and the mMDI results were 2.47 and
549 2.56, respectively, and Sabogal-Paz et al. (2020) reported mMDI values of 0.95. In the
550 study, the results of the mMDI for I-HSSF 1 were 2.22 ± 0.132 and for I-HSSF 2, it was
551 2.04 ± 0.082 , which is close to the results obtained by Andreoli and Sabogal-Paz (2020).

552 The residence time distribution for C-HSSF 1 was 293 ± 6 min and for C-HSSF 2,
553 it was 462 ± 79 min. These values are close to the theoretical residence time distribution
554 of 300 ± 8 min and 450 ± 13 min for C-HSSF1 and C-HSSF2, respectively. The
555 adjustment data for the N-CSTR model and for the models with small and large
556 dispersion of the C-HSSFs are shown in Table 3. In the C-HSSF 1 (cross-sectional area
557 = 0.028 m^2), the N-CSTR model presented the best result, requiring 22 complete mix
558 reactors in series. The largest number of reactors in series indicates a plug flow reactor
559 (Levenspiel, 1999). The C-HSSFs studied by Sabogal-Paz et al. (2020) presented $N =$
560 17 (filter cross-section of 0.0075 m^2). Those tested by Andreoli and Sabogal-Paz (2020)
561 obtained $N = 8$ and $N = 13$ (filter cross-sectional area of 0.049 m^2) and those tested by
562 Terin and Sabogal-Paz (2019) had $N = 6$ (filter cross-sectional area of 0.049 m^2). In the
563 C-HSSF 2, the piston model with great dispersion had the best result. Although C-
564 HSSF1 and C-HSSF 2 fit different models, the results of both refer to a plug flow
565 reactor.

566

567 Table 3 - Tracer test results for the C-HSSF

HSSF	Statistic	N-CSTR		Small dispersion model		High dispersion model	
		N	r ²	D, μL ⁻¹	r ²	D, μL ⁻¹	r ²
C-HSSF1	Mean	22	0.86	0.02	0.82	0.02	0.80
	Standard deviation	6	0.05	0.01	0.04	0.01	0.03
C-HSSF2	Mean	10	0.64	0.05	0.63	0.05	0.68
	Standard deviation	2	0.14	0.01	0.10	0.01	0.07

568

569 The flow in piston flow observed in I-HSSFs and C-HSSFs is good because it
 570 shows that the treatment will be homogeneous in all water fractions that enter the filter.
 571 Sabogal-Paz et al. (2020) suggested that the piston flow reactor improved the
 572 development of the biological layer and the efficiency of the treatment.

573 Figure 6 shows the comparison between experimental and simulated RTD for C-
 574 HSSFs and I-HSSFs. Results from CFD simulations are validated with experimental
 575 RTD tests (Brannock et al., 2010; Furman & Stegowski, 2011; Gharibian et al., 2020;
 576 Klusener et al., 2007; Qi et al., 2013; Talvy et al., 2011; Zhang et al., 2007). The
 577 computational RTDs of the I-HSSF1 and I-HSSF2 show that the simulations present a
 578 perfect Gaussian distribution, as expected for the simulation.

579 The form of data collection may have influenced the signal obtained
 580 experimentally, which shows the difference observed in the results of the experimental
 581 and simulated RTDs, as was found by Talvy et al. (2011) who chose to modify the
 582 simulated RTD to fit the experimental RTD. It is noteworthy that the RTDs were

583 simulated using the average velocity and constant values of the boundary condition,
584 such as porosity and permeability. When the real experiment was constructed, the layers
585 may have undergone natural and anthropic compaction, accommodating the grains
586 during assembly and feeding, and presented values different from those obtained in the
587 laboratory, which may explain the delay in the response time to the stimulus of the
588 simulated HSSFs compared to the experimental ones.

589 The Kruskal Wallis non-parametric test (5% significance level) showed that the
590 results of the three experimental tests of the RTDs and the computational test of the
591 RTD did not show statistically significant differences between each other for each of the
592 filters. The C-FLD 1, C-FLD 2, I-FLD 1 and I-FLD2 had a p-value of 0.27; 0.25; 0.01
593 and 0.38. Thus, the RTD result obtained with CFD is like the results of RTDs obtained
594 experimentally in full-scale physical models, as there was no significant difference
595 between the results.

596 The numerical simulations of the RTDs were validated with the experimental
597 tests of the RTDs, and therefore the hydrodynamics of the HSSFs were analysed. The
598 behaviour of the current lines of the C-FLDs and I-FLDs were similar and showed that
599 they are not influenced due to the geometry, the filtration rate, or the type of continuous
600 or intermittent feeding. In the C-FLDs and I-FLDs, it was found that the flow was of the
601 piston type, with no short circuit and having a dead zone close to the base. The smaller
602 the diameter, the smaller the dead zone. The C-HSSFs showed minimization adverse
603 hydraulic phenomena, as seen by the lower variation of the filter velocity profile, which
604 may reflect in the improvement of treatment efficiency due to the reduction of
605 interferences in the development of the biological layer.

606

607 **4. Conclusions**

608 The behaviour of the C-HSSFs and I-HSSFs flow lines were similar according to
609 the visual analysis of the 3D flow lines (supplementary material). In the C-HSSFs and I-
610 HSSFs, it was found that the flow was of the piston type, with no short circuit and
611 having a dead zone close to the base. The piston flow is good for the treatment because
612 the incoming water will be treated at the same time in all its portions, favouring the
613 efficiency of the HSSF.

614 The flow of the piston type is desirable for HSSFs, as it allows biological
615 treatment to occur at the beginning of the filter layer and, concomitantly, physical-
616 chemical treatments occur as the water percolates through the filter layer until the
617 treated water is obtained at the end of layers, that is, the incoming water will be treated
618 at the same time in all its portions as it percolates through the filter media.

619 The volumes of the dead zone in the HSSFs were less than 3%, and due to this
620 low value, there was no need to change the design of the HSSFs. The statistical results
621 showed that there were no significant differences in the statistical comparison of the
622 data from the experimental and simulated RTDs.

623 In view of this and considering that all filters showed flow of the piston type, the
624 choice of geometry and type of feed depends on studies of the analysis of treatment
625 efficiency considering the water quality parameters.

626

627 **Acknowledgements**

628

629 This work was supported by the Global Challenges Research Fund (GCRF) UK
630 Research and Innovation (SAFEWATER; EPSRC Grant Reference EP/P032427/1).

631

632 **Statement**

633

634 The authors hereby declare previous originality check, no conflict of interest and
635 open access to the repository of data used in this paper for scientific purposes.

636

637 **Supplementary Material**

638

639 Brazilian standards used in the physical characterization of materials and
640 experimental results are provided as supplementary material.

641

642 **References**

643 Andreoli, F. C., Sabogal-Paz, L. P., 2020. Household slow sand filter to treat
644 groundwater with microbiological risks in rural communities. *Water Research*. 186,
645 116352. <https://doi.org/10.1016/j.watres.2020.116352>.

646 ANSYS, Inc., *Ansys fluent User's Guide*, 2013. Release 15.0, ANSYS, Canonsburg,
647 United States of América.

648 Berbert, A. C., Almeida, R. A de, Rezende, R. V. de P., Lautenschlager, S. R. The use
649 of computation fluids dynamics to select outlet system configuration in biosand
650 filters. Conference: 13th IWA Specialized Conference on Small Water and
651 Wastewater Systems & 5th IWA Specialized Conference on Resources-Oriented
652 Sanitation At: Athens, September, 2016.

653 Botari, A., Di Bernardo, L., 2012. Hydrodynamic analysis of particle collection
654 efficiency: comparing downflow and upflow filtration. *Acta Scientiarum*.
655 *Technology*, v. 34, n. 2, p. 167-175. doi:10.4025/actascitechnol.v34i2.10803.

656 Brannock, M., Wang, Y., Leslie, G., 2010. Mixing characterisation of full-scale
657 membrane bioreactors: CFD modelling with experimental validation. *Water*
658 *Research* 44 (10), 3181–3191. <https://doi.org/10.1016/j.watres.2010.02.029>.

659 Bradley, I., Straub, A., Maraccini, P., Markazi, S., Nguyen, T.H., 2011. Iron oxide
660 amended biosand filters for virus removal. *Water Res* 45, 4501–4510.
661 doi:10.1016/j.watres.2011.05.045.

662 Campos, L.C., Smith, S.R., Graham, N.J.D, 2006a. Deterministic based model of slow
663 sand filtration. I: model development. *Journal of Environmental Engineering* 132(8),
664 872-886. [https://doi.org/10.1061/\(ASCE\)0733-9372\(2006\)132:8\(872\)](https://doi.org/10.1061/(ASCE)0733-9372(2006)132:8(872)).

665 Campos, L.C., Smith, S.R., Graham, N.J.D, 2006b. Deterministic based model of slow
666 sand filtration. II: model application. *Journal of Environmental Engineering* 132(8),
667 887-894. [https://doi.org/10.1061/\(ASCE\)0733-9372\(2006\)132:8\(887\)](https://doi.org/10.1061/(ASCE)0733-9372(2006)132:8(887)).

668 CAWST, 2012. Biosand Filter Construction Manual. Centre for affordable water and
669 sanitation. Canada, Calgary.

670 CAWST, 2009. Biosand filter manual: Design, construction, installation, operation and
671 maintenance, Canada, Calgary, Canada.

672 Courant, R., K. Friedrichs e H. Lewy. On the Partial Difference Equations of
673 Mathematics Physics. *IBM Journal*, n. March, p. 215-234. 1967.

674 Dadonaite, B., Ritchie, H., Roser, M., 2020. Diarrheal diseases. Published online at
675 OurWorldInData.org. (<https://ourworldindata.org/diarrheal-diseases>, accessed 19
676 May 2020).

677 Darcy, H. Les fontaines publiques de la ville de Dijon. Paris: Victor Dalmont, 1856. 647
678 p.

679 Elliott, M.A., Stauber, C.E., Koksal, F., DiGiano, F.A., Sobsey, M.D., 2008. Reductions
680 of *E. coli*, echovirus type 12 and bacteriophages in an intermittently operated

681 household-scale slow sand filter. *Water Res* 42, 2662–2670.
682 doi:10.1016/j.watres.2008.01.016.

683 Ergun, S. Fluid flow through packed columns. *Chemical Engineering Progress*, v. 48, n.
684 2, p. 89-94, 1952.

685 Forchheimer, P. Wasserbewegung durch Boden. *Zeitschrift des Vereins Deutscher*
686 *Ingenieure*, v. 45, n. 49, p. 1736-1741, 1901a.

687 Forchheimer, P. Wasserbewegung durch Boden. *Zeitschrift des Vereins Deutscher*
688 *Ingenieure*, v. 45, n. 50, p. 1781-1788, 1901b.

689 Furman, L., Stegowski, Z., 2011. CFD models of jet mixing and their validation by
690 tracer experiments. *Chemical Engineering and Processing: Process Intensification*
691 50(3), 300–304. <https://doi.org/10.1016/j.cep.2011.01.007>.

692 Gharibian, S., Hazrati, H., Rostamizadeh, M. 2020. Continuous electrooxidation of
693 Methylene Blue infilter press electrochemical flowcell: CFD simulation and RTD
694 validation. *Chemical Engineering & Processing: Process Intensification*. 150 (2020)
695 107880. <https://doi.org/10.1016/j.cep.2020.107880>.

696 Jadhav, D., Husain, M., Chavan, F.I., 2015. Biosand Water Filter: A Boon. *International*
697 *Research Journal of Engineering and Technology* 2 (3).
698 <https://www.irjet.net/archives/V2/i3/Irjet-v2i3262.pdf>

699 Kennedy, T. J., Hernandez, E. A., Morse, A. N., Anderson, T. A. 2012. Hydraulic
700 Loading Rate Effect on Removal Rates in a BioSand Filter: A Pilot Study of Three
701 Conditions. *Water, Air, & Soil Pollution*, 223 (7), 4527-4537.
702 <https://link.springer.com/article/10.1007/s11270-012-1215-4>.

703 Kikkawa, I. 2008. Modification of a Biosand Filter in the Northern Region of Ghana.
704 Master of Engineering in Civil and Environmental Engineering at the Massachusetts
705 Institute of Technology.

706 Klusener, P. A. A., Jonkers, G., During, F., Hollander, E. D., Schellekens, C. J.,
707 Ploemen, I. H. J., Othman, A., Bos, A. N. R., 2007. Horizontal cross-flow bubble
708 column reactors: CFD and validation by plant scale tracer experiments. *Chemical*
709 *Engineering Science* 62(18–20), 5495–5502.
710 <https://doi.org/10.1016/j.ces.2007.03.044>.

711 Langergraber, G., Giraldi, D., Mena, J., Meyer, D., Peña, M., Toscano, A., Brovelli, A.,
712 Korkusuz, E. A., 2009. Recent developments in numerical modelling of subsurface
713 flow constructed wetlands. *Science of The Total Environment* 407 (13), 3931–3943.
714 <https://doi.org/10.1016/j.scitotenv.2008.07.057>.

715 Levenspiel, O., 1999. *Chemical Reaction Engineering*, Industrial & Engineering
716 Chemistry Research. <https://doi.org/10.1021/ie990488g>.

717 Lynn, T.J., Wanjugi, P., Harwood, V.J., Ergas, S.J., 2013. Dynamic performance of
718 biosand filters. *J. Am. Water Works Assoc.* 105 (10), E587–E595.
719 <https://doi.org/10.5942/jawwa.2013.105.0116>.

720 Maciel, P. M. F.; Sabogal-Paz, L. P. (2020). Household slow sand filters with and
721 without water level control: continuous and intermittent flow efficiencies.
722 *Environmental Technology*, 41(8), 944-958.
723 <https://doi.org/10.1080/09593330.2018.1515988>.

724 Navier, C. L. M. H. Mémoire sur les lois du mouvement des fluides. *Mémoires de l’*
725 *Académie Royale des Sciences de l’Institut de France*, v.6, p. 389-440, 1823.

726 Qi, W. K.; Guo, Y. L.; Xue, M.; LI, Y. Y. 2013. Hydraulic analysis of an upflow sand
727 filter: Tracer experiments, mathematical model and CFD computation. *Chemical*
728 *Engineering Science*. 104, 460-472. <https://doi.org/10.1016/j.ces.2013.09.035>.

729 Robinson, R. A.; Stoke, R. H. 1955. *Electrolyte solutions*. Butterworths Publications.
730 Londres.

731 Sabogal-Paz, L. P.; Campos, L. C.; Bogush, A.; Canales, M. 2020. Household slow
732 sand filters in intermittent and continuous flows to treat water containing low mineral
733 ion concentrations and Bisphenol A. *Science of the Total Environment*. 702, 135078.
734 <https://doi.org/10.1016/j.scitotenv.2019.135078>.

735 Sobsey, M.D., Stauber, C.E., Casanova, L.M., Brown, J.M., Elliot, M.A., 2008. Point of
736 use household drinking water filtration: a practical, effective solution for providing
737 sustained access to safe drinking water in the developing World. *Environ. Sci.*
738 *Technol.* 42, 4261e4267. <https://doi.org/10.1021/es702746n>.

739 Stokes, G. G. On the theories of internal friction of fluids in motion and of the
740 equilibrium and motion of elastic solids. *Transactions of the Cambridge*
741 *Philosophical Society*, v. 8, p. 207-305, 1845.

742 Stropky, D.; Pougatch, K.; Nowak, P.; Salcudean, M.; Pagoria, P.; Gartshore, I.; Yuan,
743 J. 2007. RTD (residence time distribution) predictions in large mechanically aerated
744 lagoons. *Water Science & Technology*. 55, 11, 29-36.
745 <https://doi.org/10.2166/wst.2007.346>.

746 Talvy, S.; Debaste, F.; Martinelli, L.; Chauveheid, E.; HAUT, B. 2011. Development of
747 a tool, using CFD, for the assessment of the disinfection process by ozonation in
748 industrial scale drinking water treatment plants. *Chemical Engineering Science*. 66,
749 3185-3194. <https://doi.org/10.1016/j.ces.2011.02.039>.

750 Tchobanoglous, G., Burton, F.L., Stensel, H.D., 2003. *Wastewater Engineering:*
751 *Treatment and Reuse*. McGraw-Hill Higher Education, New York.

752 Terin, U. C.; Sabogal-Paz, L. P. 2019. *Microcystis aeruginosa* and microcystin-LR
753 removal by household slow sand filters operating in continuous and intermittent
754 flows. *Water Research*. 150, 29-39. <https://doi.org/10.1016/j.watres.2018.11.055>.

755 United Nations Environment Programme (UNEP). 2016. A Snapshot of the World's
756 Water Quality: Towards a global assessment. Unites Nations Environment
757 Programme, Nairóbi, Quénia.

758 YOUNG-ROJANSCHI, C.; MADRAMOOTOO, C. Comparing the performance of
759 biosand filters operated with multiday residence periods. Journal of Water Supply:
760 Research and Technology-Aqua. 64, 157. 2014a.
761 <https://doi.org/10.2166/aqua.2014.027>.

762 YOUNG-ROJANSCHI, C; MADRAMOOTOO, C. Intermittent Versus Continuous
763 Operation of Biosand Filters. Water Research v. 45. 1-10 p., 2014b

764 Zhang, L., Pan, Q., Rempel, G.L., 2007. Residence time distribution in a multistage
765 agitated contactor with Newtonian fluids: CFD prediction and experimental
766 validation. Industrial and Engineering Chemistry Research 46 (11), 3538-3546.
767 <https://doi.org/10.1021/ie060567+>.

768

Supplementary Material

769

770

771 Table S1 – Standards used in the physical characterization of materials

Materials	Granulometry	CU ¹	Specific mass	Voids index	Permeability
Fine sand	ABNT NBR 11799/2016				
Coarse sand		ABNT	ABNT	ABNT	ABNT NBR 13292/1995
Fine gravel	ABNT NM 248/2001	NBR 6502/1995	NBR 6458/2016	NM 45/2006	
Coarse gravel					

772 Notes: ¹ Coefficient of uniformity (CU).

773

774 Table S2 – Specific mass, voids index and material of permeability

Materials	Specific mass (g.cm ⁻³)	Voids index	Permeability (cm.s ⁻¹)
Fine sand	2.66	0.40	1.97 x 10 ⁻²
Coarse sand	2.64	0.46	1.06 10 ⁻¹
Fine gravel	2.63	0.43	4.75
Coarse gravel	2.64	0.39	8.28

775

776 Table S3 - Characteristics of the porous media of the filters

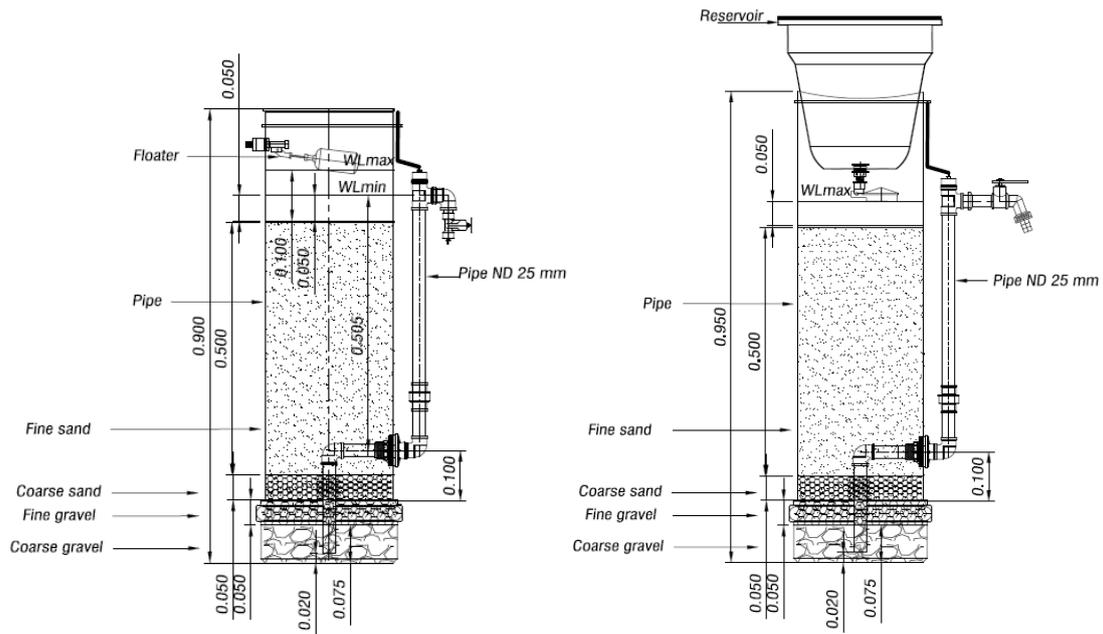
Filter materials	D _p (m)	ε ()	α (m ²)	1/α (m ⁻²)	C ₂ (m ⁻¹)
Fine sand	2.70 x10 ⁻⁴	0.40	8.64 x10 ⁻¹¹	1.16 x10 ¹⁰	1.21 x10 ⁵
Coarse sand	1.30 x10 ⁻³	0.46	3.76 x10 ⁻⁹	2.66 x10 ⁸	1.49 x10 ⁴
Fine gravel	4.00 x10 ⁻³	0.43	2.61 x10 ⁻⁸	3.83 x10 ⁷	6.27 x10 ³
Coarse gravel	1.04 x10 ⁻²	0.39	1.15 x10 ⁻⁷	8.70 x10 ⁶	3.46 x10 ³

777 Notes: 1/α: viscous resistance, C₂: inertial resistance, D_p: average particle diameter, ε:

778 voids index

779

780



781

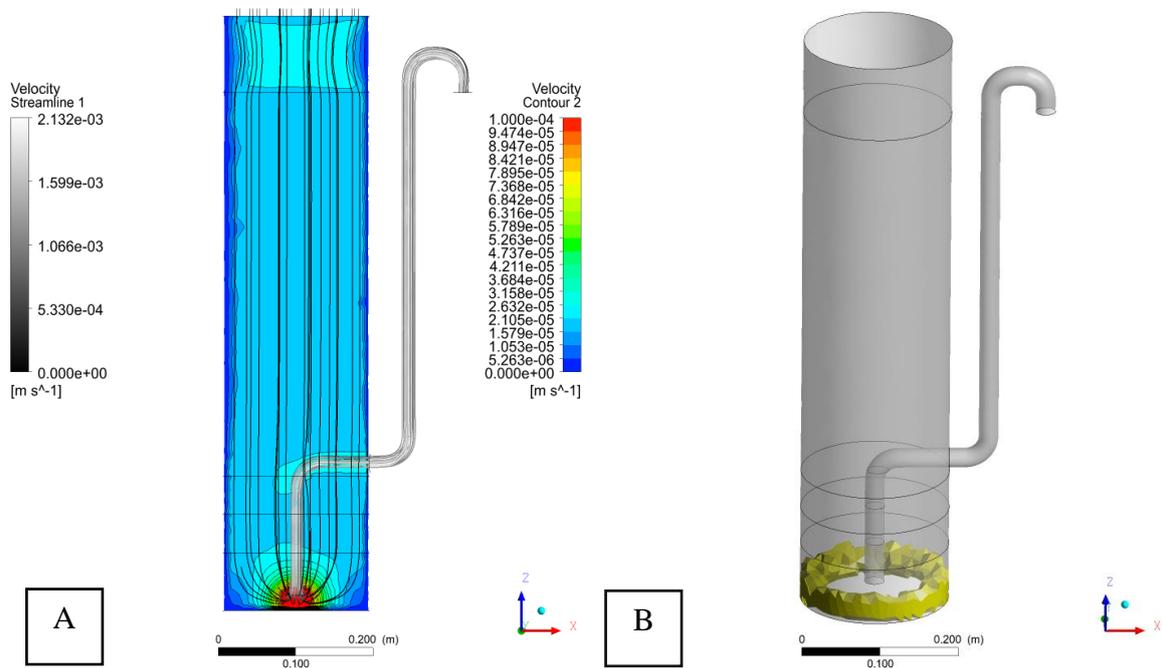
782 Continuous flow (C-HSSF)

Intermittent flow (I-HSSF)

783

Figure S1. Scheme of the studied HSSFs (units in meters)

784

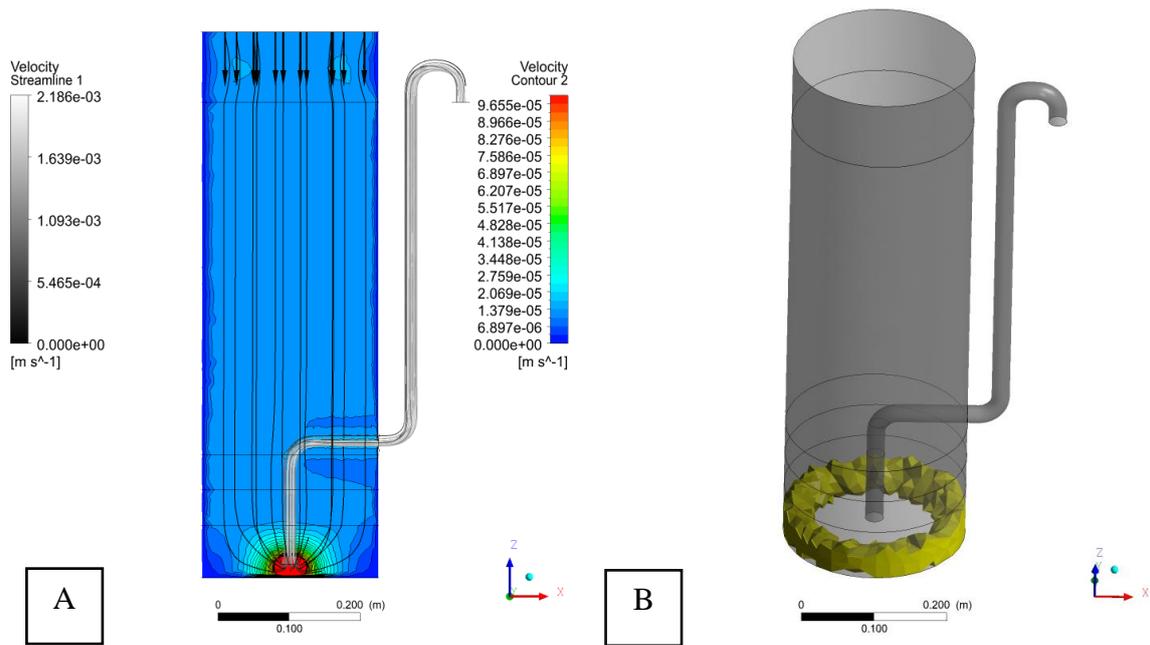


785 Figure S2 – Dead zone in C-HSSF1 (A - Current lines and velocity contour plane; B -

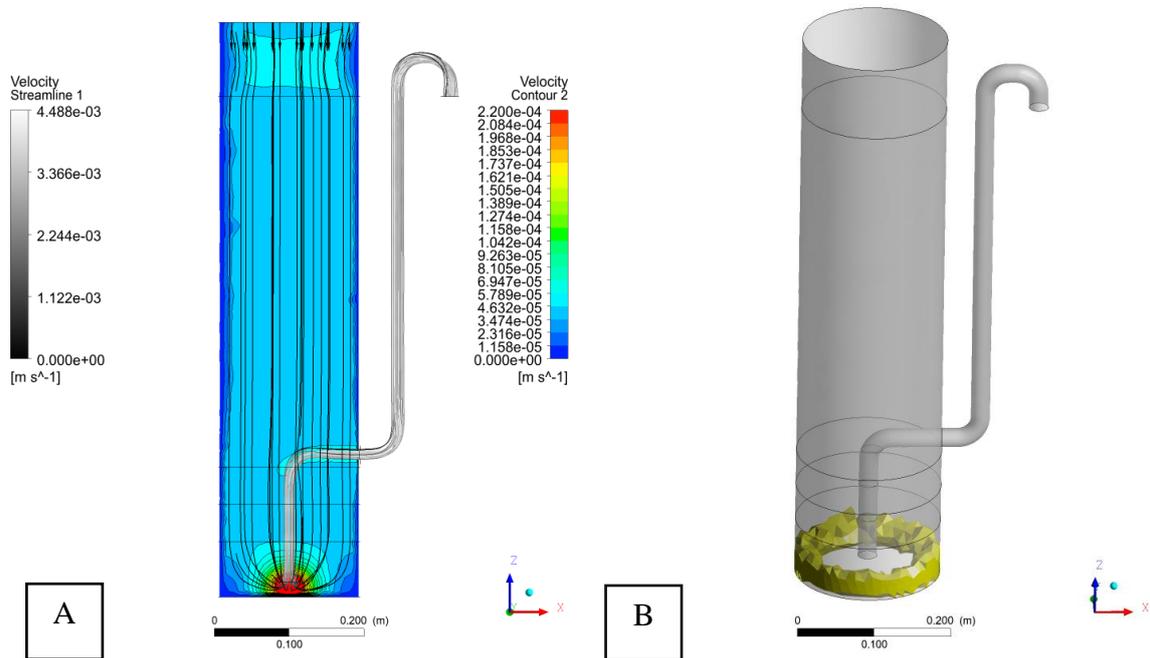
786

Dead zone volume in yellow).

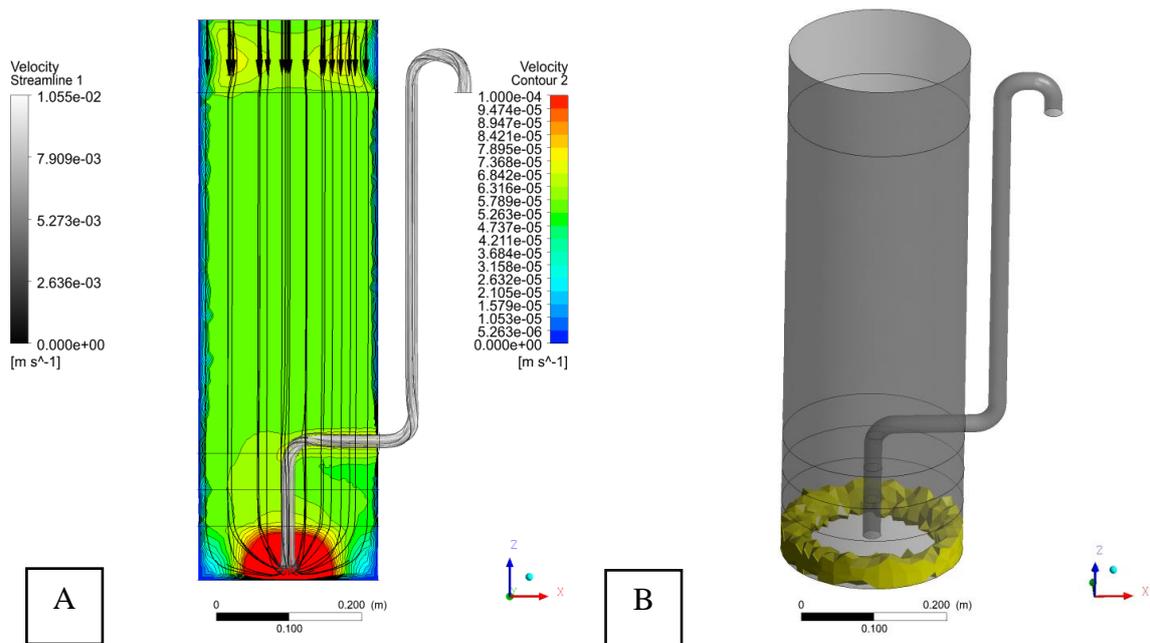
787



788 Figure S3 – Dead zone in C-HSSF2 (A - Current lines and velocity contour plane; B -
 789 Dead zone volume in yellow).



790 Figure S4 – Dead zone in I-HSSF1 (A - Current lines and velocity contour plane; B -
 791 Dead zone volume in yellow).



792 Figure S5 – Dead zone on I-HSSF2 (A - Current lines and velocity contour plane; B -
 793 Dead zone volume in yellow).

794

795 **References**

796 ABNT NBR 6458. Grãos de pedregulho retidos na peneira de abertura 4,8 mm —
 797 Determinação da massa específica, da massa específica aparente e da absorção de água.
 798 [Gravel grains retained in the 4.8 mm opening sieve - Determination of specific gravity,
 799 apparent specific gravity and water absorption]. Brazil. Rio de Janeiro, 2016.
 800 Associação Brasileira de Normas Técnicas - ABNT.

801 ABNT NBR 6502. Rochas e solos. [Rocks and soils]. Brazil. Rio de Janeiro, 1995.
 802 Associação Brasileira de Normas Técnicas - ABNT.

803 ABNT NBR 11799. Material filtrante — Areia, antracito e pedregulho —Especificação.
 804 [Filter material - Sand, anthracite and gravel - Specification]. Brazil. Rio de Janeiro,
 805 2016. Associação Brasileira de Normas Técnicas - ABNT.

806 ABNT NBR 13292. Solo – Determinação do coeficiente de permeabilidade de solos
807 granulares à carga constante. [Determination of the permeability coefficient of granular
808 soils at constant load]. Brazil. Rio de Janeiro, 1995. Associação Brasileira de Normas
809 Técnicas - ABNT.

810 ABNT NM 45. Agregados – Determinação da massa unitária e do volume de vazios.
811 [Aggregates - Determination of unit mass and void volume]. Brazil. Rio de Janeiro,
812 2006. Associação Brasileira de Normas Técnicas - ABNT. Norma Mercosul – NM.

813 ABNT NM 248. Agregados – Determinação da composição granulométrica.
814 [Aggregates - Determination of particle size composition]. Brazil. Rio de Janeiro, 2001.
815 Associação Brasileira de Normas Técnicas - ABNT. Norma Mercosul – NM.